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ARTICLE

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Using the multivariate Hawkes process to study interactions between multiple species from camera trap data

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Abstract

Interspecific interactions can influence species' activity and movement patterns. In particular, species may avoid or attract each other through reactive responses in space and/or time. However, data and methods to study such reactive interactions have remained scarce and were generally limited to two interacting species. At this time, the deployment of camera traps opens new opportunities but adapted statistical techniques are still required to analyze interaction patterns with such data. We present the multivariate Hawkes process (MHP) and show how it can be used to analyze interactions between several species using camera trap data. Hawkes processes use flexible pairwise interaction functions, allowing us to consider asymmetries and variations over time when depicting reactive temporal interactions. After describing the theoretical foundations of the MHP, we outline how its framework can be used to study interspecific interactions with camera trap data. We design a simulation study to evaluate the performance of the MHP and of another existing method to infer interactions from camera trap-like data. We also use the MHP to infer reactive interactions from real camera trap data for five species from

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South African savannas (impala *Aepyceros melampus*, greater kudu *Tragelaphus strepsiceros*, lion *Panthera leo*, blue wildebeest *Connochaetes taurinus* and Burchell's zebra *Equus quagga burchelli*). The simulation study shows that the MHP can be used as a tool to benchmark other methods of interspecific interaction inference and that this model can reliably infer interactions when enough data are considered. The analysis of real data highlights evidence of predator avoidance by prey and herbivore–herbivore attraction. Lastly, we present the advantages and limits of the MHP and discuss how it can be improved to infer attraction/avoidance patterns more reliably. As camera traps are increasingly used, the multivariate Hawkes process provides a promising framework to decipher the complexity of interactions structuring ecological communities.

KEYWORDS
African savanna, camera trap, interaction network, interspecific interactions, multivariate Hawkes process, reactive response, Snapshot Safari, spatio-temporal interactions

INTRODUCTION

Interspecific interactions affect many aspects of ecological communities. For instance, they influence ecosystem services (Valiente-Banuet et al., 2015), species assembly via biotic filtering (Ovaskainen et al., 2017) and the behavior of interacting species. In particular, interactions are one of the factors that structure the way in which animal species move in the landscape and adjust their habitat choices or activity times (Palmer et al., 2022). Mobile animals can respond to interactions by avoiding or seeking proximity with individuals of other species, depending on the positive or negative outcome of the interactions. For instance, prey can avoid their predators (Say-Sallaz et al., 2019), competing species can avoid each other (Cornhill et al., 2022; Searle et al., 2021), or herbivores can forage together to reduce predation risk or increase access to preferred foraging resources (Beaudrot et al., 2020). In this paper, we will use the term "interaction" to refer to the attraction or avoidance of a species by another one, even though "interaction" also refers to the underlying process of the attraction/avoidance pattern.

These interactions (as defined above) can occur in space and/or time, at different scales. Species can adjust their space use in response to the expected distribution of other species (proactive spatial interaction; Palmer et al., 2022). Species can also alter their daily activity patterns (e.g., Karanth et al., 2017) in response to other species (proactive temporal interaction). However, some species could also exhibit a reactive response to the presence of other species, that is, change their behavior in response to the actual presence of a species sometime before at a given location (e.g., Karanth et al., 2017; Parsons et al., 2016). This type of response could be mediated, for instance, by olfactory (Cornhill & Kerley, 2020; Kuijper et al., 2014) or auditory cues (Hettena et al., 2014). Investigating these reactive interactions is particularly promising as it allows us to identify fine-grained patterns that could be missed by approaches aggregating data in space or in time (Cusack et al., 2017; Frey et al., 2017; Parsons et al., 2022).

Investigating such fine-scale responses is very challenging, as it requires an intensive sampling effort to monitor multiple species in space and time. In this context, camera traps open new opportunities to study the spatial and temporal activities of multiple species (Caravaggi et al., 2017). Camera trap arrays allow the collection of multiple species occurrences and, therefore, the continuous monitoring of entire communities for large areas in time (Pardo et al., 2021). Hence, camera traps can produce massive amounts of data and offer new possibilities to study interactions between several species at multiple scales. Moreover, they are relatively cheap and easier to set up than classical fieldwork survey techniques (e.g., transects), especially for rare or elusive species or in remote areas. As camera traps become more affordable and automated species identification methods from pictures are being developed with deep learning, camera trap data (and other passive sensor data) will probably become more abundant in the future (Caravaggi et al., 2017).

With camera trap data, interspecific interactions are mostly studied at a broad spatial or temporal scale. To do this, data are often aggregated so that either the spatial or the temporal aspect is completely ignored. There are two main approaches for this purpose: comparing species' daily activities patterns (Ridout & Linkie, 2009) or spatial

occupancy patterns (e.g., with the multispecies occupancy model of Rota et al., 2016). Such methods provide a measure of the proactive attraction or avoidance strategy, with species adapting their space or time use in anticipation of other species' presence or absence (Palmer et al., 2022). However, other approaches have combined spatial and temporal aspects to infer reactive attraction/avoidance strategies (frequently called spatio-temporal interactions in the literature; Karanth et al., 2017; Murphy et al., 2021; Niedballa et al., 2019; Prat-Guitart et al., 2020). Most methods using camera trap data quantify only the temporal aspect of reactive interactions; therefore, we will call the inferred patterns *reactive temporal interactions*. Most are based on the computation of time intervals between the detections of two species at a given place (e.g., Harmsen et al., 2009); here, we call this family of methods *interevent times methods*. The distribution of time intervals can then be contrasted according to the order of appearance of species (Parsons et al., 2016; Prat-Guitart et al., 2020) or summarized by a statistic that is compared with values obtained under a null model (usually data permutation; Cusack et al., 2017; Galindo-Aguilar et al., 2022; Karanth et al., 2017; Murphy et al., 2021). For a comparison of different approaches to infer reactive temporal avoidance with time interval measures, see Niedballa et al. (2019). Other more recent approaches use point processes, which allow us to integrate temporal dependence in a model-based framework (Kellner et al., 2022; Schliep et al., 2018).

Although all methods described above are useful to study reactive interactions, they usually focus on pairs of species and can therefore be unsuitable for studying complex interaction networks. For instance, these methods can identify spurious interactions between two species if other species are involved in the interaction network but not considered in the analysis. Moreover, they summarize the effect of a species on another one by a single value (e.g., the median of the time interval; Karanth et al., 2017), thus ignoring the multiscale and possibly time-dependent changes in the attraction/avoidance patterns (but see Cusack et al., 2017).

In this paper, we propose the multivariate Hawkes process (MHP) (Hawkes, 1971; Lambert et al., 2018) as a modeling framework to infer reactive interactions between multiple species from passive sensors such as camera traps. Hawkes processes belong to the family of point processes that allow the analysis of species capture events in continuous time, thus avoiding any data aggregation procedure. In Hawkes processes, species' interactions are modeled as pairwise interaction functions that depend on the time elapsed between species detections. The MHP used in this article is generative and offers the possibility to simulate occurrence data, given parameters specification. It also comes with an inference procedure that allows the adjustment of the pairwise interaction functions from observed data. It deals properly with indirect effects caused by species interaction chains, thus minimizing the risk of inferring spurious interactions. We believe that this model is a useful conceptual framework that is well suited for assessing reactive temporal interactions between species. We first present the Hawkes process and how it can be used to analyze camera trap data. Then, we describe the MHP used in this article, which was developed by Lambert et al. (2018) and implemented in the R package *UnitEvents* (Albert et al., 2021). We then show how this model can be used to simulate data to evaluate the performance of statistical methods or to infer interactions from camera trap data. We also apply the MHP on real camera trap data from the Snapshot Safari monitoring program (Pardo et al., 2021) to infer reactive temporal interactions between five mammal species. Finally, we discuss the usefulness of the MHP and the perspectives on how to develop this model further.

MATERIALS AND METHODS

All analyses were performed using R statistical software (v4.3.0; R Core Team, 2023) and the code and data (Nicvert et al., 2023) are available at https://doi.org/10.6084/m9.figshare.24552157.v5.

Model: The multivariate Hawkes process (MHP)

Hawkes processes are a family of point processes used to describe dependencies between punctual events. These processes belong to the class of self-exciting point processes for which the probability of occurrence at time t depends on the occurrences of the previous events. The first Hawkes process was introduced in 1971 by Alan G. Hawkes (Hawkes, 1971). Originally applied to model aftershocks following earthquakes (e.g., Ogata, 1988), Hawkes processes have been applied in various fields (Reinhart, 2018) for instance to model crime recurrence in cities (Mohler et al., 2018), the evolution of prices on the stock market (Hawkes, 2018) or the transmission of action potentials in a network of neurons (Reynaud-Bouret et al., 2013). The theoretical properties of Hawkes processes have also been thoroughly studied, and numerous extensions have been proposed.

Throughout this article, we define an occurrence as the detection of an individual at a camera at a given time,

and we do not take imperfect detection into account. To describe the model, we consider data on the occurrences of S species collected on C cameras. In our framework, the data collected on C cameras are seen as C independent realizations of the MHP. Let T_m^{li} denote the m-th instant of punctual occurrence for species i at camera l. Let N_i^l be the total number of occurrences for species i at camera l. We model the occurrence times $\left(T_m^{li}\right)_{m=1\ldots N_i^l, i=1\ldots S, l=1\ldots C}$ as C realizations of an MHP.

To model punctual occurrences, point processes use a latent intensity function, which is a measure of the rate at which events occur in time. When modeling species occurrences from camera trap data, the intensity for a given species represents the rate at which this species occurs at a camera. For species i, the intensity $\lambda_i^l(t)$ at camera l is formally defined as (Daley & Vere-Jones, 2003):

$$\lambda_i^l(t) = \lim_{\delta \to 0} \frac{P\left\{n^{li}_{]t,t+\delta]} > 0\right\}}{\delta}, \tag{1}$$

where $n^{li}_{]t,t+\delta]}$ is the number of points occurring between times t and $t+\delta$ for species i at camera l and δ is an infinitesimally small amount of time. Informally, the intensity of a point process multiplied by a small amount of time can be viewed as the probability that there will be at least one point occurring around time t.

In this work, we used the R package *UnitEvents* (Albert et al., 2021) available from https://sourcesup.renater.fr/frs/?group_id=3267, to simulate and infer MHPs. *UnitEvents* is only available on Linux and Mac OS. However, in the code and data repository for the article (Nicvert et al., 2023), we provide a Dockerfile allowing us to run the analyses from any operating system (including Windows).

UnitEvents implements the MHP described in Lambert et al. (2018). In this framework, the intensity of species i seen on camera l for a Hawkes process with S interacting species is written as:

$$\lambda_i^l(t) = \left(\nu_i + \sum_{j=1}^{S} \sum_{m|T_m^{lj}<t} f_{j\to i}\left(t - T_m^{lj}\right)\right)_+, \tag{2}$$

where $\lambda_i^l(t)$ represents the intensity for species i (as defined above) at camera l. ν_i is a positive parameter, the background rate: it represents the basal intensity of species i (in time^{-1}, e.g., day^{-1}) unrelated to previous occurrences. For instance, ν_i would be low for a rare species and higher for a common species.

$f_{j\to i}$ is the interaction function that represents the influence of an occurrence of species j on species i as a function of time delay: positive values of $f_{j\to i}$ represent an attraction of species i by species j, negative values represent a repulsion and null values independence. In the case $j=i$, the function $f_{i\to i}$ represents the interaction between individuals of the same species i. In that case, we will call $f_{i\to i}$ the auto-interaction function: it could reflect for instance the fact that some species are solitary or gregarious. $f_{j\to i}$ are defined as piecewise constant functions with K time bins of equal length δ:

$$f_{j\to i} = \sum_{k=1}^{K} \alpha_{j\to i}^k \mathbf{1}_{](k-1)\delta,k\delta]}, \tag{3}$$

where $\mathbf{1}_{](k-1)\delta,k\delta]}$ denotes the indicator function between delays $(k-1)\delta$ and $k\delta$. The K coefficients $\alpha_{j\to i}^k$ represent the average number of occurrences of species i gained (if positive) or suppressed (if negative) by an occurrence of species j in the k-th interval after this occurrence of species j.

In this framework, $f_{j\to i}$ can take negative values, thus allowing modeling the repulsive effect of species j on species i. As the intensity λ_i^l must be positive by definition, Equation (2) includes a positive part $(\cdot)_+$. However, for mathematical reasons, in the following developments we will assume that the negative values of $f_{j\to i}$ are never too strong so that the intensity never becomes negative, and the positive part is not needed. To enforce this assumption, the repulsion terms can only be as strong as the other terms making up the total intensity.

Figure 1 illustrates a realization of an MHP with five species (measured at a single camera) simulated with *UnitEvents*. In this example, some species attract each other (see the interaction network in Figure 1a) with the same decreasing discrete exponential interaction function with $K=12$ time bins of width $\delta=4$ h (Figure 1b). The background rate is the same for all species and is fixed at 0.2 occurrences day^{-1}. The right panel in Figure 1c shows the simulated species occurrences and associated intensities over time. When nothing happens, the intensity is fixed at the background rate. When an attracting species occurs, the intensity of the attracted species peaks, making an occurrence more likely. For instance, each occurrence of species s_1 gives rise to a peak in the intensity of s_2. Moreover, when several attracting events occur, the interaction functions add up, which makes the occurrence of the target species even more likely.

Model inference

The inference procedure implemented in the *UnitEvents* package is a fast and scalable LASSO-penalized (least absolute shrinkage and selection operator) least-squares criterion. It allows the estimation of a single MHP from C

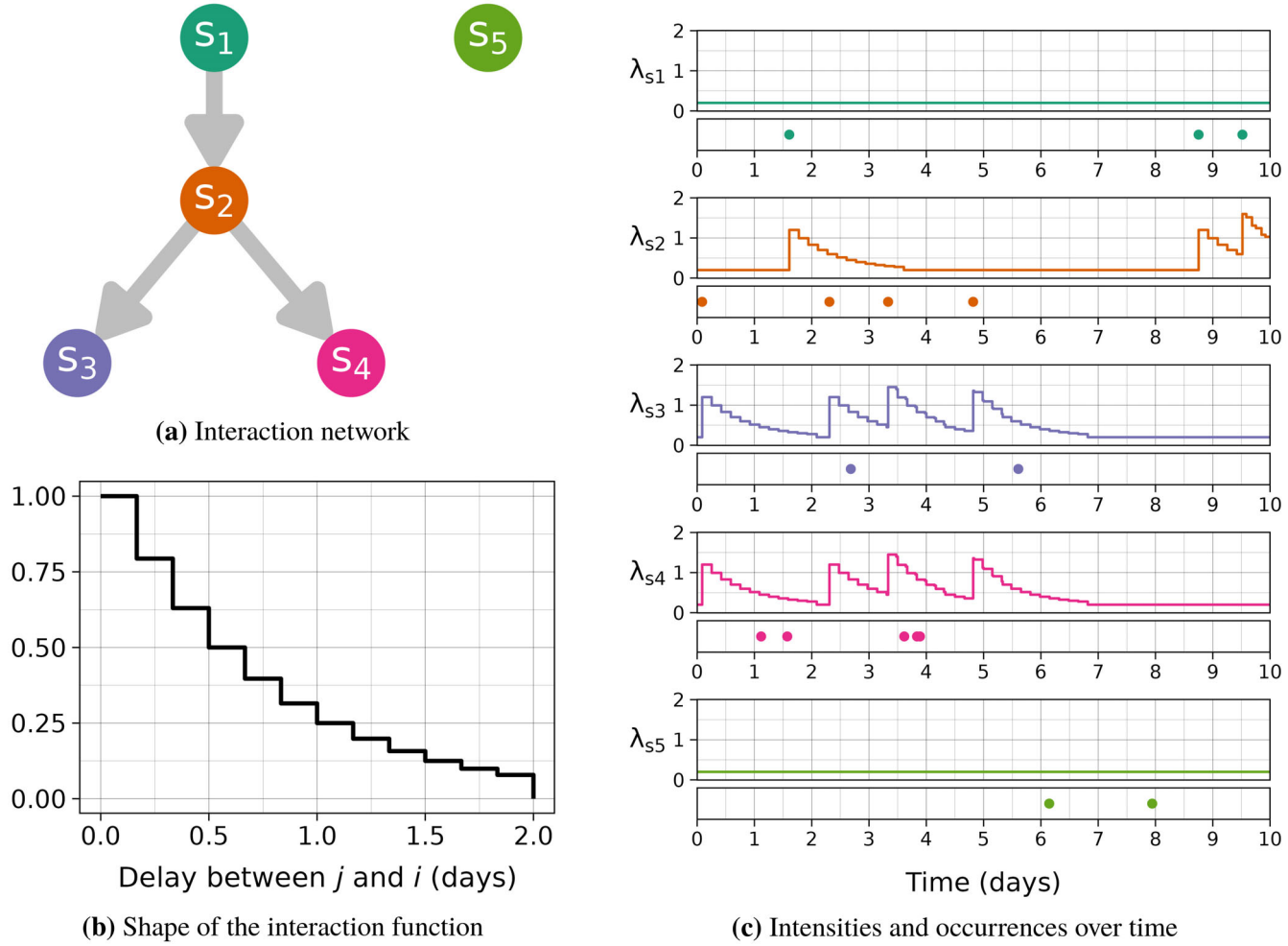

FIGURE 1 Example of a realization of a multivariate Hawkes process. (a) Shows the interaction network between five species (each arrow represents a non-null interaction function). In this example, all autointeraction functions $f_{i\to i}$ are null. (b) Shows the shape of the interaction functions ($K = 12$ time bins of width $\delta = 4$ h) corresponding to arrows in the interaction network (a). (c) Shows a realization of the Hawkes process with the interaction network and the interaction functions shown in (a) and (b). For this simulation, the background rate was set to 0.2 occurrences day^{-1} for all species. For each species i, the above panel shows the intensity λ_i and the bottom panel shows the species occurrences. Each time an attracting species occurs, the intensity for the attracted species peaks and then decreases as dictated by the interaction function shape.

realizations (in our setting, this corresponds to C cameras).

Let $(\boldsymbol{\beta}_1, \ldots, \boldsymbol{\beta}_S)$ denote the parameters of interest for each species $i = 1, \ldots, S$. Each $\boldsymbol{\beta}_i$ is a vector of size $1 + SK$ containing the background rate of species i (ν_i) and the parameters of the interaction functions targeted to this species i for the S species and the K bins: $\boldsymbol{\beta}_i = \left(\nu_i, \left(\alpha_{j\to i}^k\right)_{j=1\ldots S, k=1\ldots K}\right)$. Each $\boldsymbol{\beta}_i$ is estimated as:

$$\widehat{\boldsymbol{\beta}_i} = \arg\min_{\boldsymbol{\beta}_i} \mathrm{LASSO}(\boldsymbol{\beta}_i) \quad \text{where}$$

$$\mathrm{LASSO}(\boldsymbol{\beta}_i) = \underbrace{-2\sum_{l=1}^{C} \boldsymbol{b}_i^{l\,T}\boldsymbol{\beta}_i + \boldsymbol{\beta}_i^T \sum_{l=1}^{C} \boldsymbol{G}^l \boldsymbol{\beta}_i}_{\text{least-squares}} + \underbrace{2\boldsymbol{d}_i^T |\boldsymbol{\beta}_i|}_{\text{penalization}}, \quad (4)$$

where ^{T} denotes transposition and $|\boldsymbol{\beta}_i|$ is the vector containing the absolute values of the coordinates of $\boldsymbol{\beta}_i$. $\boldsymbol{b}_i^l$ is an observable vector of size $1 + SK$. If camera l is active between times α_l and η_l, then:

$$\boldsymbol{b}_i^l = \left(N_i^l, \left(\int_{\alpha_l}^{\eta_l} n_{[t-k\delta, t-(k-1)\delta[}^{lj} \mathrm{d}n_t^{li}\right)_{j=1\ldots S, k=1\ldots K}\right). \quad (5)$$

Its first value is the total count of species i observed on camera l. The other values represent the total occurrence counts of the species j observed in the k-th bin before the occurrences of species i at camera l. $\boldsymbol{G}^l$ is also an observable matrix defined as:

$$\boldsymbol{G}^l = \int_{\alpha_l}^{\eta_l} \boldsymbol{c}_t^l \boldsymbol{c}_t^{l\,T} \mathrm{d}t, \tag{6}$$

where $\boldsymbol{c}_t^l$ is a vector of size $1+SK$ defined as $\boldsymbol{c}_t^l = (1, (n^{lj}_{[t-k\delta, t-(k-1)\delta]})_{j=1\ldots S, k=1\ldots K})$. Its first value is 1 and other values represent the occurrence counts of species j occurring on camera l in the k-th bin before time t.

The term $2\boldsymbol{d}_i^T \,|\, \boldsymbol{\beta}_i \,|$ of Equation (4) corresponds to the LASSO penalization: it can make some parameter values shrink to zero and thus avoid overparameterization. The strength of this LASSO penalization is controlled by the weights vectors $\boldsymbol{d}_i$, which are computed from the data and tuned by a unique user-chosen parameter γ (equation derived from Lambert et al. (2018) adapted from Hansen et al. (2015)):

$$\boldsymbol{d}_i = \sqrt{2\gamma \log\left(S+S^2K\right) \sum_{l=1}^{C} \int_{\alpha_l}^{\eta_l} \boldsymbol{c}_t^{l\,2} \mathrm{d}n_t^{li}} + \frac{\gamma \log\left(S+S^2K\right)}{3} \max_{l=1\ldots C} \left(\sup_{t \in [\alpha_l, \eta_l]} |\boldsymbol{c}_t^l| \right). \tag{7}$$

The choice of a suitable value for γ is crucial for model selection, because γ ensures that only relevant nonzero parameters are kept in the model. However, choosing a good value for γ is difficult: it has been evaluated by simulations in Lambert et al. (2018) and Hansen et al. (2015), and we proceeded similarly in this article.

In the current implementation of *UnitEvents*, three flavors of the LASSO penalization are available. We chose the "Bernstein Vanishing LASSO" (BVL), where the penalization in Equation (4) is first applied to discard weak interaction parameters. Then, the estimates of the remaining non-null parameters are obtained by minimizing the least-squares criterion. Lastly, an additional step is introduced to remove parameters smaller than a data-computed threshold (see Lambert et al., 2018, for details and justification).

In the implementation of *UnitEvents*, the bins width δ and the number of bins K for the interaction functions are fixed by the user, who also needs to choose a value of γ *a priori*. The other parameters (interaction functions coefficients $\alpha^k_{j\to i}$ and background rates ν_i) are fitted as described before.

Simulation study

We generated camera trap-like data under the MHP and used these simulated data to (1) evaluate the performance of a method and (2) tune the penalization parameter for inference on real data.

Simulation parameters

For these two objectives, we conducted two sets of simulations in the same conditions. We considered an interaction network with five species $s_{i=1\ldots5}$ where s_1 attracts s_2 and s_2 attracts s_3 and s_4 (network from Figure 1a). This network represents a difficult case as an inference method should detect direct interactions, but not spurious indirect interactions (e.g., $s_1 \to s_3$) and identify that species s_5 is not interacting with others. In this simulation, we define the true interactions by decreasing exponential functions to 2 days:

$$f(t) = \begin{cases} \alpha \exp\left(-\dfrac{\ln(2)}{0.5} t\right) & \text{if } t < 2 \\ 0 & \text{if } t \geq 2 \end{cases}, \tag{8}$$

where α is the interaction strength. The half-life of this function is the denominator of the decrease rate, so that this function will reach half of its initial value at $t = 0.5$ day. The interaction strength α for the true model varied from 0.01 to 1 day^{-1}. Here, the interaction strength represents the maximum intensity of the pairwise interaction function for $t = 0$. An analogous interaction function is shown in Figure 1b with $\alpha = 1$ and with discrete bins. The background rate was fixed at 0.1 day^{-1} for all species.

The simulated trapping length varied from 20 to 500 trapping days for each camera over 25 cameras (making up to 12,500 trapping days in total). For each condition, 30 different data sets were generated to evaluate the variability of the inference.

We evaluated the performance of the inference by computing the true positive and true negative rates. The true positive rate is the proportion of inferred nonzero interactions over the count of true nonzero interactions. The true negative rate is the proportion of inferred null interactions over the count of true null interactions.

Evaluating a method to infer reactive temporal interactions

We illustrated how synthetic data generated with the MHP can be used to evaluate the performance of a method to infer interspecific interactions, considering the interevent times method of Murphy et al. (2021). We applied the method described by Murphy et al. (2021) on simulated data (simulation settings are described in *Simulation parameters*). This method consists of computing the median time between directed pairwise species occurrences (excluding pairs from the same species) for observed and randomly permuted data (999 permutations). The permutation procedure involved randomly

changing the cameras' labels of species occurrences (for details see Murphy et al., 2021). Finally, the statistical significance of interactions was estimated by comparing the median time for observed and permuted data. We used a significance threshold of 5% with a Holm correction for multiple testing.

Choice of the penalization parameter

To choose the best penalization parameter γ in the context of interaction inference, we used a simulation approach (simulation settings are described in *Simulation parameters*). We inferred MHPs with different values of γ (between 0.3 and 1) from the simulated datasets. For the inference parameters, we chose $K = 12$ bins of width $\delta = 4$ h (2 days in total, corresponding to the length of the simulated interaction functions). Then, we defined any inferred interaction function as null if all bins were zero over the function's support, and nonnull if at least one bin was not null.

Application: Analysis of interactions between five species in the African savanna

We used the MHP to infer interaction functions between five species of the southern African savanna: impala *Aepyceros melampus*, greater kudu *Tragelaphus strepsiceros*, lion *Panthera leo*, blue wildebeest *Connochaetes taurinus*, and Burchell's zebra *Equus quagga burchelli*.

Data collection

Camera trap data were collected as part of the long-term Snapshot Safari monitoring program (Pardo et al., 2021). Snapshot Safari is a network of camera trap grids set up in more than 30 locations in southern Africa. The camera trap design consists of grids of 5 km^2 in each location, in which cameras were fixed at ~50 cm high. Cameras were automatically triggered by motion or heat using passive infrared sensors. Each camera was programmed to take a series of three images within 1–5 s of each other by day, and only one image by night to minimize disturbance occasioned by white flash. For this analysis, we focused on six camera trap grids in the savanna biome in northern South Africa: the Associated Private Nature Reserves (around Kruger National Park), Kruger National Park, Madikwe Game Reserve, Pilanesberg National Park, Somkhanda Game Reserve, and Venetia Limpopo Nature Reserve (see Figure 2).

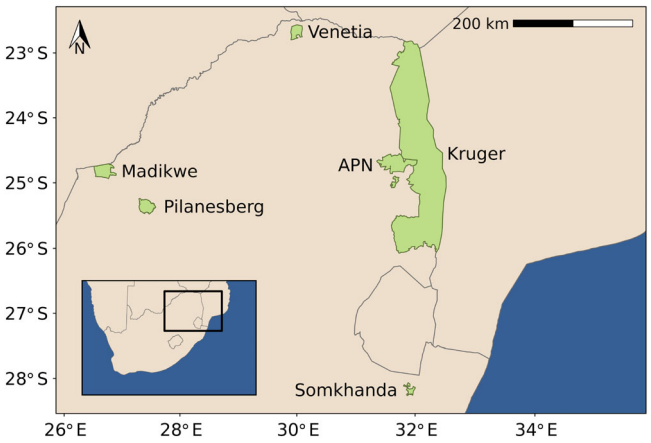

FIGURE 2 Study sites. Six protected areas were surveyed with camera traps for this study: the Associated Private Nature Reserves (APN), Kruger National Park, Madikwe Game Reserve, Pilanesberg National Park, Somkhanda Game Reserve, and Venetia Limpopo Nature Reserve. Data by OpenStreetMap contributors under ODbL license (https://www.openstreetmap.org/copyright).

Data preprocessing

Pictures were classified by citizen science using the Zooniverse platform (www.zooniverse.org), where pictures were available online and annotated by more than 150,000 volunteers (see Pardo et al., 2021, for more details).

For this analysis, we filtered out cameras where capture events were too rare (less than two pictures in total or fewer than one picture every 30 days on average). We did not filter for independence between occurrences of the same species. However, because the Hawkes model does not allow two capture events to occur simultaneously, if two or more individuals of different species were seen on the same capture event, their occurrence time was randomly shifted from 1 min in advance to 1 min later. For multiple individuals of the same species seen simultaneously, the occurrences of the individuals were counted as a single event (i.e., an occurrence corresponds to an individual or a group of individuals of a given species).

After the filtering procedure, 72,703 occurrence events (corresponding to 70,409 unique pictures) were collected on 179 cameras in total. Cameras were active during 503 ± 224 (SD) days on average (minimum: 19 days, maximum: 851 days), amounting to 90,176 trapping days on all cameras. All pictures were taken between June 2017 and November 2019.

Parameters inference

We inferred the parameters of a MHP using interaction functions defined by $K = 6$ bins of $\delta = 6$ h (36 h in total). This parametrization should allow us to capture the

dynamics of reactive temporal interactions with enough granularity while keeping a relatively low number of parameters to estimate to allow reliable inference. Using the results of the simulation study (see *Choice of the penalization parameter*), we decided to set the value of the penalization parameter γ to 0.5.

RESULTS

Simulation study

Evaluating a method to infer reactive temporal interactions

We used data simulated under the MHP to evaluate the method of Murphy et al. (2021). As expected, the ability to detect interactions (true positive rate) increases with the strength of the interactions (Figure 3). Provided the interaction strength α is big enough (at least 0.1 day^{-1}), the ability to detect interactions increases with the number of trapping days, which indicates that a significant sampling effort is required to infer interactions from camera trap data (at least 300 trapping days for 25 cameras when the interaction strength is above 0.2 day^{-1}). More surprisingly, when the interaction strength is high (at least 0.5 day^{-1}), the true negative rate decreases with increasing sampling effort. This indicates that the method wrongly detects interactions between noninteracting species. Additional investigations (Appendix S1: Section S1) show that these errors mainly concern the detection of spurious indirect interactions between species involved in interaction chains (e.g., $s_1 \rightarrow s_3$).

Choice of the penalization parameter

The simulation study to find suitable values for the penalization parameter γ led to the results shown in Figure 4. Unsurprisingly, the ability to detect true interactions (true positive rate) increases with the number of trapping days and the strength of interactions. When the penalization is too low ($\gamma = 0.3$; top row), the model tends to identify interactions between noninteracting species (reducing the true negative rate) but this problem vanishes when the sampling effort increases. Conversely, a high penalization ($\gamma = 1$; bottom row), moderately improves the true negative rate, but more importantly dramatically hampers the ability to detect non-null interactions for small interaction strengths. A value of $\gamma = 0.5$ seems to be a good compromise allowing the efficient detection of true interactions when their strength is not too small (at least 0.1 day^{-1}) but avoiding the identification of false interactions. It gives good results especially when the sampling lasts more than 400 trapping days per camera. Hence, we decided to use a penalization parameter of $\gamma = 0.5$ to infer the parameters of an MHP from real data (see *Analysis of real data*). Lastly, supplementary analyses show that the spurious interactions are randomly distributed and not biased toward indirect interactions as with the interevent time method (Appendix S1: Section S1).

Analysis of real data

We fitted a MHP using the occurrence data of five species (impala, greater kudu, lion, blue wildebeest, and Burchell's

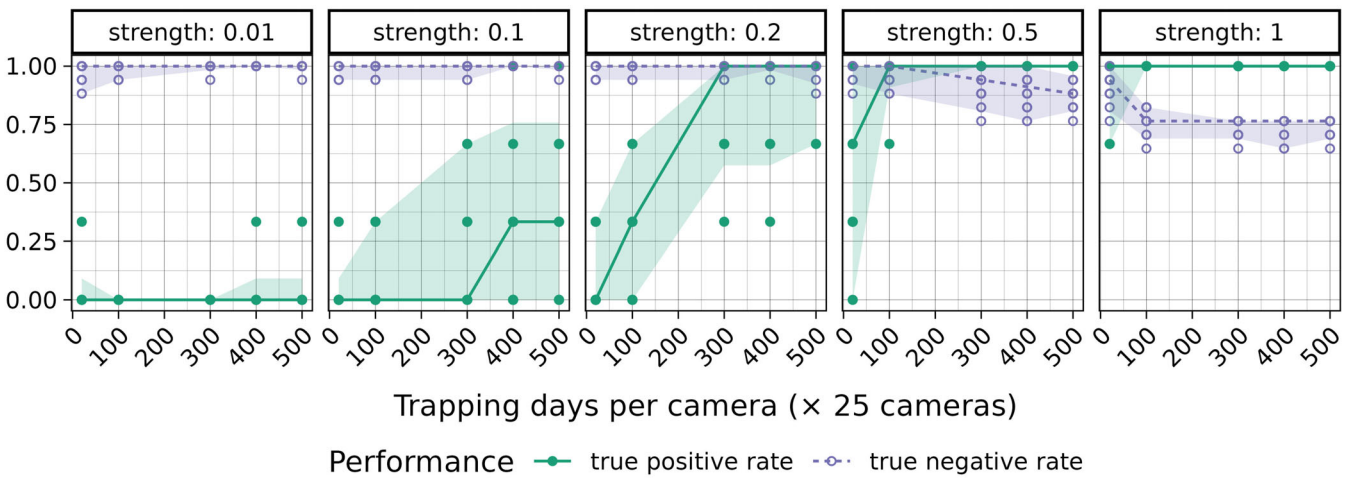

FIGURE 3 Evaluation of an interevent times method (Murphy et al., 2021). Panels represent different interaction strengths (maximum value of the interaction function). The *x*-axis represents the sampling length and the *y*-axis represents the performance: true positive rate (full dots, continuous line) or true negative rate (circles, dashed line). Points indicate values for the 30 repetitions, lines joins the medians, and the colored area represents the 2.5th and 97.5th percentiles.

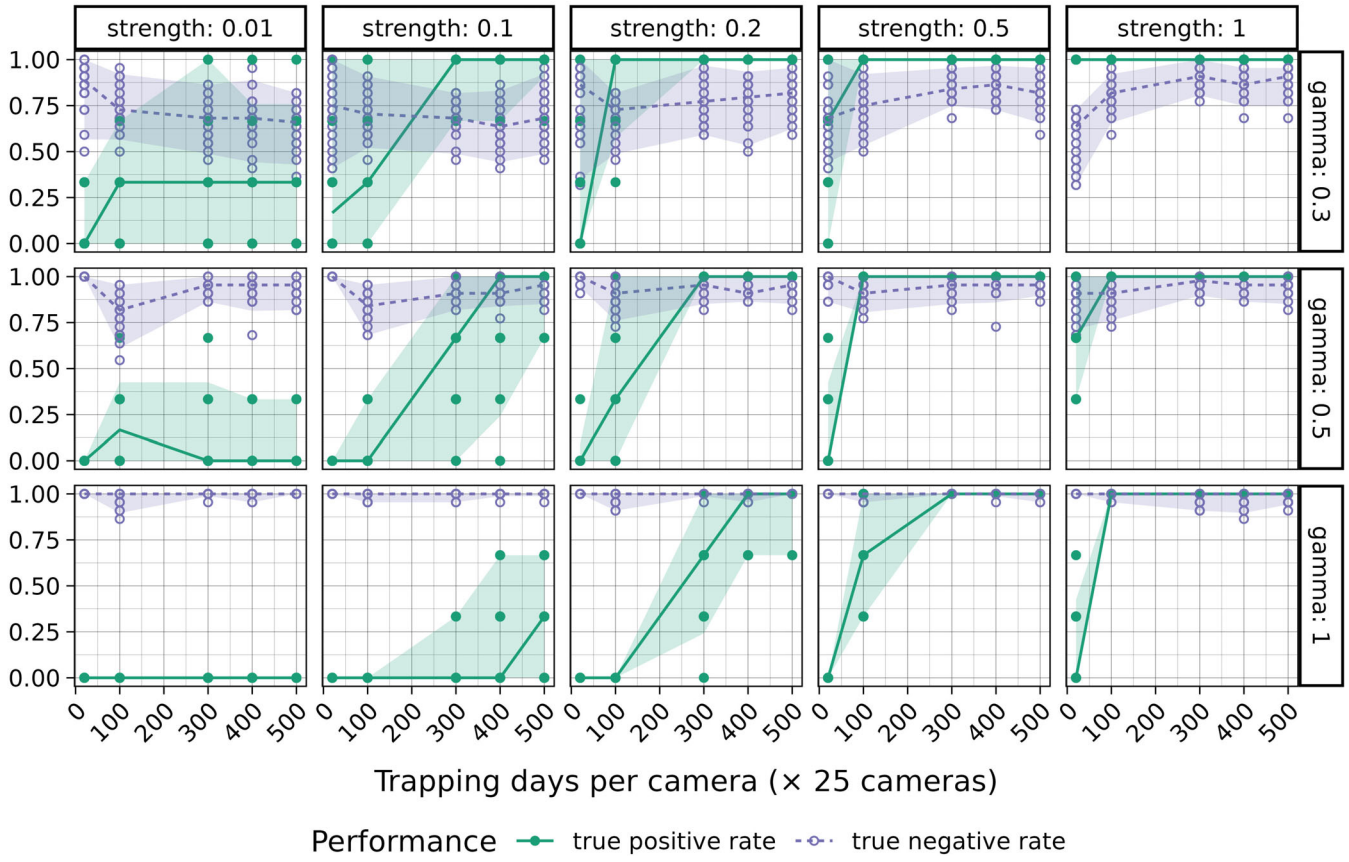

FIGURE 4 Performance of the inference with the multivariate Hawkes model. In columns, the interaction strength (maximum value of the interaction function). In rows, the different values of the penalization parameter γ. The *x*-axis represents the sampling length and the *y*-axis represents the performance (true positive rate or true negative rate). Lines, points and colors have the same meaning as in Figure 3.

zebra) collected with camera traps. Adjusting the model only took a few seconds on a personal computer. The resulting interaction functions are shown in Figure 5 and the inferred background rates are in Appendix S1: Section S2.

Background rates represent the basal intensity for each species, independently of the others. They vary greatly between species, with impala having a much higher background rate than other species (impala: 0.212 day^{-1}; zebra: 0.040 day^{-1}; kudu: 0.035 day^{-1}; wildebeest: 0.022 day^{-1}, and lion: 0.003 day^{-1}). As expected, they are strongly related to the total occurrence count of each species.

Regarding the interaction functions, the inferred parameters highlight a strong auto-attraction for the first bin (0–6 h), varying between 1.5 and 2.25 day^{-1} depending on the species. Regarding the cross-species interaction functions, many herbivores are attracted to each other. Impalas follow or avoid kudus (depending on the delay), wildebeests and zebras; zebras follow impalas, kudus, and wildebeests; wildebeests mainly follow zebras. Other interactions between herbivores are negligible. These herbivore–herbivore interactions are composed of a short-term attraction (during the first 6 h after an occurrence) and of a medium-term attraction (12 to 36 h after an occurrence) except impalas that are not attracted by zebras in the short term. Additionally, impalas seem to avoid kudus 6 to 12 and 30 to 36 h after an occurrence. We notice that these interactions are asymmetrical (impalas and zebras follow other species much more than they are followed). Regarding prey–predator interactions, lions do not follow or avoid any other species. Zebra and impala seem to avoid lions in the next 6 h following an occurrence of this predator. Finally, the inferred interactions are relatively robust to a change in bin width as we show in Appendix S1: Section S4, where we performed the inference on the same dataset with different bin widths (3 and 9 h).

DISCUSSION

It is now well established that identifying the signature of interspecific interactions from species occurrence data is generally difficult (Blanchet et al., 2020; Popovic et al., 2019). However, camera trap data provide additional

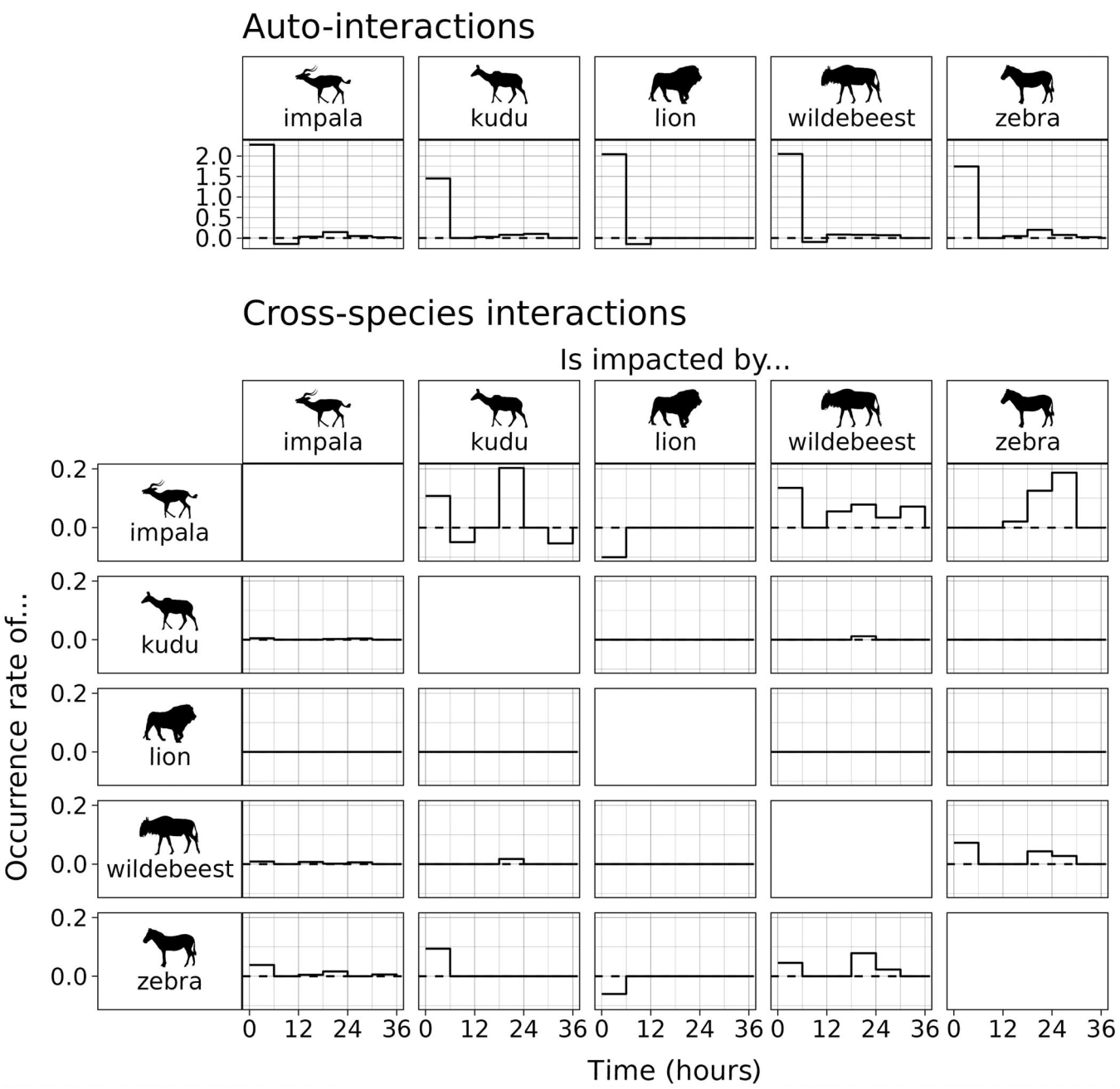

FIGURE 5 Inference of interactions from real data using the multivariate Hawkes model. The top plot shows the auto-interaction functions (between occurrences of the same species). The bottom plot shows cross-species interactions, where the intensity of species in rows is affected by species in columns. The horizontal dashed line represents zero. Note that the *y*-axis scale is different between autointeractions and cross-species interactions. Silhouette images from PhyloPic by Lukasiniho (wildebeest), Margot Michaud (lion), Robert Hering (kudu), Zimices (zebra), and an unknown author (impala).

information (time and order of occurrence of species) that can help to relate occurrence patterns to underlying interspecific interactions. In this context, the Hawkes model provides a new theoretical framework to analyze species occurrences sampled in continuous time using camera traps. This model aims to predict the probability of occurrence of a given species at a given time taking into account the previous occurrences for several species. By considering the exact time at which species occur, this model provides a detailed picture of species reactive temporal interactions under the form of interaction functions (here, the term "interaction" refers to the attraction/repulsion pattern). These functions allow a multiscale description of interactions as they characterize how the interaction strength varies with time, contrary to other methods that provide a single measure of attraction/

avoidance. Moreover, these functions are directed: the inferred interactions can be asymmetrical, as expected for ecological interactions. The toolbox associated with this model offers the possibility to generate data and design simulation studies or to infer parameters from real data.

We used the MHP to generate camera trap-like datasets with different properties (sampling effort, strength of interactions) and showed how these simulated data can be used to evaluate the performance of a method or to tune inference parameters. Both simulation studies demonstrate that camera trap data can be used to detect reactive temporal interactions between species but this requires a substantial sampling effort, especially when the strengths of interactions are low. In our simulation setup (five species, background rates of 0.1 day^{-1} and only attractions), results suggest that at least 1 year of sampling with 25 cameras is required to obtain reliable inference, and this holds only if the interaction is strong enough (interactions of strength 0.01 day^{-1} are not reliably detected in our simulations). These requirements would probably be higher if more species were considered, especially for rare species (smaller background rate). Hence, we agree with Schliep et al. (2018) that more data are needed to estimate reliable reactive interaction patterns than to estimate species occupancy. In this context, the MHP provides a powerful simulation tool to design and assess the quality of sampling protocols in camera trap studies by adopting a virtual ecologist approach (Zurell et al., 2010). The simulation study also highlights the limits of methods focusing on pairs of species to analyze interactions between multiple species. By focusing only on two species at a time, these approaches are not able to disentangle direct interactions from indirect effects due to other species in interaction chains (Appendix S1: Section S1). Moreover, the correction for multiple testing we applied in our study was not sufficient to eliminate these spurious interactions, and we can assume that this issue is more important in the literature when no correction is considered. As a consequence, interevent times methods tend to overestimate the number of interactions, especially when their strength is high or the sampling effort increases. However, such spurious interactions were inferred only when we simulated quite strong interactions, and more investigations would be needed to estimate the range of interaction strengths we can expect in natural conditions. By contrast, the Hawkes process used here is multivariate by nature, so it works on all species simultaneously and thus allows the identification of interactions between two species conditionally to the other species, similar to graphical models in the context of co-occurrence analysis (Popovic et al., 2019). This modeling approach thus provides a better picture of the interaction network of the whole community.

The real dataset analysis shows how the MHP can be used to infer reactive interactions between five mammal species from the African savanna. In our example, because we defined an occurrence as the presence of an individual or a group of individuals, the values of the interaction functions represent the number of individuals or groups of individuals that are attracted/repulsed by other occurrences, and the typical group size to consider depends of the species. We identified strong auto-attractions for all species but also attractions between different herbivores and avoidance of lions by two herbivore species (impala and zebra). Whereas it could be tempting to interpret these results as behavioral responses of species to an underlying interaction (e.g., avoidance in response to predation), the Hawkes model only characterizes attraction/avoidance patterns and particular care should be taken when interpreting these results, especially because no covariates were included in this analysis. We discuss these different interpretations of the observed patterns in terms of ecological processes below and we make suggestions to improve the MHP to untangle the different hypotheses.

We identified auto-attractions for all species, indicating that the occurrence of a given species increases the probability of having another occurrence of the same species at the same place. This could be due to the same individual lingering in front of the camera, especially because no independence filter was applied (although cameras are configured to pause for 1 min between trigger events), or this could reflect sociality among individuals, as an individual or a group may attract other individuals for gregarious species. This could also stem from habitat selection processes, so that numerous subsequent occurrences could be observed at cameras located in species' preferred habitats. Lastly, circadian rhythms impose physiological constraints on the activity times of each species and thus could increase their occurrence rate at certain times of the day. When they are not taken into account, as is the case here, circadian rhythms could affect the interaction functions in the short term (0–6 h) and also induce a 24-h periodicity in the interaction functions. This issue is clearly illustrated using simulated data (see Appendix S1: Section S3) and could partly explain the short-term (0–6 h) auto-attraction, and probably most of the weak auto-attraction observed at ~24 h for impala, kudu, wildebeest and zebra (Figure 5).

Regarding the cross-species interactions, we observed attraction patterns between some herbivores, which could be explained by four mechanisms. First, temporal niche convergence could induce attraction between species when they are active at the same time of day and if they also share the same location. In our example, the four herbivore species are diurnal with crepuscular activity peaks. However, if the apparent attraction was due to shared

circadian rhythms, we would probably observe a symmetry of interaction functions (i.e., if $s_1 \rightarrow s_2$ is not null, $s_2 \rightarrow s_1$ is also not null) as the order of appearance of species at a camera during the activity time would be random. This is not always the case in the example depicted here, for instance between zebra and kudu. Second, species sharing the same kind of preferred environment might show apparent attraction. However, as for the temporal niche, this spatial niche should induce a symmetrical interaction pattern. Third, the apparent attraction between species could be due to the mixed-species grouping strategy, whereby some species forage together in mixed groups. Such groups are thought to mitigate predation risk and/or improve access to resources (Beaudrot et al., 2020). Moreover, when it comes to predation risk, in addition to the dilution effect (a simple number game), there is a possible benefit to being associated with more vulnerable species (Fitzgibbon, 1990). This implies a directionality in the choices of association, an asymmetry well captured by the MHP. In our analysis, some species are attracted by others in the first hours following an occurrence (impala follows kudu and wildebeest; zebra follows wildebeest, kudu and impala; and wildebeest follows zebra). Interestingly, these associations have been described in the literature (Meise et al., 2019; Pays et al., 2014; Schmitt et al., 2014). Finally, another mechanism that could explain interactions between herbivore species is grazing succession (Bell, 1971), which describes a strategy by which species sequentially use the same grazing area: less selective species come first (nonruminants and species with higher body mass), followed by more selective species (smaller ruminants). In our results, some herbivore species are attracted with a delay (impala following zebra, kudu, and wildebeest; wildebeest following zebra; zebra following wildebeest and impala). Impala following other (bigger) species and wildebeest following the nonruminant zebra are compatible with the grazing succession theory (Bell, 1971). However, the temporal scale of this potential grazing succession occurs at a temporal scale much shorter than the one classically described (McNaughton, 1976, 1985).

Regarding the apparent avoidance of lions by zebras and impalas, here again this could stem from temporal niche divergence (lion is a nocturnal species whereas impala and zebra are diurnal). This apparent repulsion could also reflect a strategy of impala and zebra to minimize predation risk by reactively avoiding lions, that is, responding to actual cues of lion presence (olfactory or auditory cues for instance) at a fine spatio-temporal scale, as documented for zebras (Courbin et al., 2016).

As discussed with the real dataset analysis, a major challenge remains linking attraction/repulsion patterns identified by the MHP to underlying ecological processes. To date, the implementation used in this paper cannot include covariates to model variations in species' background occurrence rates. This calls for two major improvements: first, we could include temporal covariates to account for the variation of species occurrence rate through the day according to their diel cycle. Second, we could include environmental covariates to account for species habitat preferences across the landscape. Works such as Fujita et al. (2018) for temporal covariates or Carstensen et al. (2010) for temporal and environmental covariates could be helpful in this perspective. Further developments include accounting for imperfect detection by camera traps, which is known to be an important issue (Burton et al., 2015). In this regard, Kellner et al. (2022) recently developed an occupancy model with a detection process occurring in continuous time with a Markov-modulated Poisson process, and a similar approach could be envisioned with the MHP.

Here, we inferred an MHP from camera trap data, but this modeling approach could be extended to other types of passive sensors collecting occurrence data in continuous time (e.g., microphones, hydrophones) that are increasingly used to monitor biodiversity. In particular, using a spatially explicit extension of the Hawkes process (first described by Ogata, 1998, in the context of earthquake occurrences) could be especially suited to include a spatial dependency between camera traps or to analyze GPS collar data and estimate interaction functions in time and space.

The Hawkes process could also be used for other applications than estimating interspecific interactions, for instance to study behavioral synchrony within a group (e.g., Pays et al., 2012) or to infer animal social networks from occurrence data (e.g., Jacoby et al., 2016).

Even if more developments are required to improve ecological inference, we contend that the MHP and other point-process methods offer an adapted theoretical framework for the analysis of time-continuous occurrence data while contributing to an explanation of interactions among herbivores and between herbivores and predators.

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CONFLICT OF INTEREST STATEMENT

The authors declare no conflicts of interest.

DATA AVAILABILITY STATEMENT

Data and code (Nicvert et al., 2023) are available on Figshare at https://doi.org/10.6084/m9.figshare.24552157.v5.

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REFERENCES

Albert, M., Y. Bouret, J. Chevallier, M. Fromont, F. Grammont, T. Laloe, C. Mascart, et al. 2021. "UnitEvents: Unitary Events Method with Delayed Coincidence Count (MTGAUE or Permutation Method) and Bernstein Lasso Method for Hawkes Processes." https://sourcesup.renater.fr/frs/?group_id=3267.

Beaudrot, L., M. S. Palmer, T. M. Anderson, and C. Packer. 2020. "Mixed-Species Groups of Serengeti Grazers: A Test of the Stress Gradient Hypothesis." *Ecology* 101: e03163.

Bell, R. H. 1971. "A Grazing Ecosystem in the Serengeti." *Scientific American* 225: 86–93.

Blanchet, F. G., K. Cazelles, and D. Gravel. 2020. "Co-Occurrence Is Not Evidence of Ecological Interactions." *Ecology Letters* 23: 1050–63.

Burton, A. C., E. Neilson, D. Moreira, A. Ladle, R. Steenweg, J. T. Fisher, E. Bayne, and S. Boutin. 2015. "Wildlife Camera Trapping: A Review and Recommendations for Linking Surveys to Ecological Processes." *Journal of Applied Ecology* 52: 675–685.

Caravaggi, A., P. B. Banks, A. C. Burton, C. M. V. Finlay, P. M. Haswell, M. W. Hayward, M. J. Rowcliffe, and M. D. Wood. 2017. "A Review of Camera Trapping for Conservation Behaviour Research." *Remote Sensing in Ecology and Conservation* 3: 109–122.

Carstensen, L., A. Sandelin, O. Winther, and N. R. Hansen. 2010. "Multivariate Hawkes Process Models of the Occurrence of Regulatory Elements." *BMC Bioinformatics* 11: 456.

Cornhill, K. L., C. Kelly, and G. I. H. Kerley. 2022. "Lion Reintroduction Demonstrates that Resident Cheetah Have a Spatially Reactive Response to Lion." *African Journal of Ecology* 60: 1–12.

Cornhill, K. L., and G. I. H. Kerley. 2020. "Cheetah Communication at Scent-Marking Sites Can be Inhibited or Delayed by Predators." *Behavioral Ecology and Sociobiology* 74: 21.

Courbin, N., A. J. Loveridge, D. W. Macdonald, H. Fritz, M. Valeix, E. T. Makuwe, and S. Chamaillé-Jammes. 2016. "Reactive Responses of Zebras to Lion Encounters Shape their Predator-Prey Space Game at Large Scale." *Oikos* 125: 829–838.

Cusack, J. J., A. J. Dickman, M. Kalyahe, J. M. Rowcliffe, C. Carbone, D. W. MacDonald, and T. Coulson. 2017. "Revealing Kleptoparasitic and Predatory Tendencies in an African Mammal Community Using Camera Traps: A Comparison of Spatiotemporal Approaches." *Oikos* 126: 812–822.

Daley, D. J., and D. Vere-Jones. 2003. *An Introduction to the Theory of Point Processes*, 2nd ed. New York: Springer.

Fitzgibbon, C. D. 1990. "Mixed-Species Grouping in Thomson's and Grant's Gazelles: The Antipredator Benefits." *Animal Behaviour* 39: 1116–26.

Frey, S., J. T. Fisher, A. C. Burton, and J. P. Volpe. 2017. "Investigating Animal Activity Patterns and Temporal Niche Partitioning Using Camera-Trap Data: Challenges and Opportunities." *Remote Sensing in Ecology and Conservation* 3: 123–132.

Fujita, K., A. Medvedev, S. Koyama, R. Lambiotte, and S. Shinomoto. 2018. "Identifying Exogenous and Endogenous Activity in Social Media." *Physical Review E* 98: 052304.

Galindo-Aguilar, R. E., B. C. Luna-Olivera, M. Ramírez-Ibáñez, and M. C. Lavariega. 2022. "Spatiotemporal co-Occurrence of Predators and Prey in a Neotropical Mammal Community in Southern Mexico." *Journal of Tropical Ecology* 38: 285–294.

Hansen, N. R., P. Reynaud-Bouret, and V. Rivoirard. 2015. "Lasso and Probabilistic Inequalities for Multivariate Point Processes." *Bernoulli* 21: 83–143.

Harmsen, B. J., R. J. Foster, S. C. Silver, L. E. T. Ostro, and C. P. Doncaster. 2009. "Spatial and Temporal Interactions of Sympatric Jaguars (*Panthera onca*) and Pumas (*Puma concolor*) in a Neotropical Forest." *Journal of Mammalogy* 90: 612–620.

Hawkes, A. G. 1971. "Spectra of some Self-Exciting and Mutually Exciting Point Processes." *Biometrika* 58: 83–90.

Hawkes, A. G. 2018. "Hawkes Processes and their Applications to Finance: A Review." *Quantitative Finance* 18: 193–98.

Hettena, A. M., N. Munoz, and D. T. Blumstein. 2014. "Prey Responses to Predator's Sounds: A Review and Empirical Study." *Ethology* 120: 427–452.

Jacoby, D. M. P., Y. P. Papastamatiou, and R. Freeman. 2016. "Inferring Animal Social Networks and Leadership: Applications for Passive Monitoring Arrays." *Journal of the Royal Society Interface* 13: 20160676.

Karanth, K. U., A. Srivathsa, D. Vasudev, M. Puri, R. Parameshwaran, and N. S. Kumar. 2017. "Spatio-Temporal Interactions Facilitate Large Carnivore Sympatry across a Resource Gradient." *Proceedings of the Royal Society B: Biological Sciences* 284: 20161860.

Kellner, K. F., A. W. Parsons, R. Kays, J. J. Millspaugh, and C. T. Rota. 2022. "A Two-Species Occupancy Model with a Continuous-Time Detection Process Reveals Spatial and Temporal Interactions." *Journal of Agricultural, Biological and Environmental Statistics* 27: 321–338.

Kuijper, D. P. J., M. Verwijmeren, M. Churski, A. Zbyryt, K. Schmidt, B. Jędrzejewska, and C. Smit. 2014. "What Cues Do Ungulates Use to Assess Predation Risk in Dense Temperate Forests?" *PLoS One* 9: e84607.

Lambert, R. C., C. Tuleau-Malot, T. Bessaih, V. Rivoirard, Y. Bouret, N. Leresche, and P. Reynaud-Bouret. 2018. "Reconstructing the Functional Connectivity of Multiple Spike Trains Using Hawkes Models." *Journal of Neuroscience Methods* 297: 9–21.

McNaughton, S. J. 1976. "Serengeti Migratory Wildebeest: Facilitation of Energy Flow by Grazing." *Science* 191: 92–94.

McNaughton, S. J. 1985. "Ecology of a Grazing Ecosystem: The Serengeti." *Ecological Monographs* 55: 259–294.

Meise, K., D. W. Franks, and J. Bro-Jørgensen. 2019. "Using Social Network Analysis of Mixed-Species Groups in African Savannah Herbivores to Assess how Community Structure Responds to Environmental Change." *Philosophical Transactions of the Royal Society B* 374: 20190009.

Mohler, G., J. Carter, and R. Raje. 2018. "Improving Social Harm Indices with a Modulated Hawkes Process." *International Journal of Forecasting* 34: 431–39.

Murphy, A., D. R. Diefenbach, M. Ternent, M. Lovallo, and D. Miller. 2021. "Threading the Needle: How Humans Influence Predator-Prey Spatiotemporal Interactions in a Multiple-Predator System." *Journal of Animal Ecology* 90: 2377–90.

Nicvert, L., S. Donnet, M. Keith, M. Peel, M. J. Somers, L. H. Swanepoel, and J. A. Venter. 2023. "Code and Data for: Using the Multivariate Hawkes Process to Study Interactions between Multiple Species from Camera Trap Data." https://doi.org/10.6084/m9.figshare.24552157.v5.

Niedballa, J., A. Wilting, R. Sollmann, H. Hofer, and A. Courtiol. 2019. "Assessing Analytical Methods for Detecting Spatiotemporal Interactions between Species from Camera Trapping Data." *Remote Sensing in Ecology and Conservation* 5: 272–285.

Ogata, Y. 1988. "Statistical Models for Earthquake Occurrences and Residual Analysis for Point Processes." *Journal of the American Statistical Association* 83: 9–27.

Ogata, Y. 1998. "Space-Time Point-Process Models for Earthquake Occurrences." *Annals of the Institute of Statistical Mathematics* 50: 379–402.

Ovaskainen, O., G. Tikhonov, A. Norberg, F. Guillaume Blanchet, L. Duan, D. Dunson, T. Roslin, and N. Abrego. 2017. "How to Make More out of Community Data? A Conceptual Framework and its Implementation as Models and Software." *Ecology Letters* 20: 561–576.

Palmer, M. S., K. M. Gaynor, J. A. Becker, J. O. Abraham, M. A. Mumma, and R. M. Pringle. 2022. "Dynamic Landscapes of Fear: Understanding Spatiotemporal Risk." *Trends in Ecology & Evolution* 37: 911–925.

Pardo, L. E., S. P. Bombaci, S. Huebner, M. J. Somers, H. Fritz, C. Downs, A. Guthmann, et al. 2021. "Snapshot Safari: A Large-Scale Collaborative to Monitor Africa's Remarkable Biodiversity." *South African Journal of Science* 117: 1–4.

Parsons, A. W., C. Bland, T. Forrester, M. C. Baker-Whatton, S. G. Schuttler, W. J. McShea, R. Costello, and R. Kays. 2016. "The Ecological Impact of Humans and Dogs on Wildlife in Protected Areas in Eastern North America." *Biological Conservation* 203: 75–88.

Parsons, A. W., K. F. Kellner, C. T. Rota, S. G. Schuttler, J. J. Millspaugh, and R. W. Kays. 2022. "The Effect of Urbanization on Spatiotemporal Interactions between Gray Foxes and Coyotes." *Ecosphere* 13: e3993.

Pays, O., A. Ekori, and H. Fritz. 2014. "On the Advantages of Mixed-Species Groups: Impalas Adjust their Vigilance when Associated with Larger Prey Herbivores." *Ethology* 120: 1207–16.

Pays, O., E. Sirot, and H. Fritz. 2012. "Collective Vigilance in the Greater Kudu: Towards a Better Understanding of Synchronization Patterns." *Ethology* 118: 1–9.

Popovic, G. C., D. I. Warton, F. J. Thomson, F. K. Hui, and A. T. Moles. 2019. "Untangling Direct Species Associations from Indirect Mediator Species Effects with Graphical Models." *Methods in Ecology and Evolution* 10: 1571–83.

Prat-Guitart, M., D. P. Onorato, J. E. Hines, and M. K. Oli. 2020. "Spatiotemporal Pattern of Interactions between an Apex Predator and Sympatric Species." *Journal of Mammalogy* 101: 1279–88.

R Core Team. 2023. *R: A Language and Environment for Statistical Computing*. Vienna: R Foundation for Statistical Computing. https://www.R-project.org/.

Reinhart, A. 2018. "A Review of Self-Exciting Spatio-Temporal Point Processes and their Applications." *Statistical Science* 33: 299–318.

Reynaud-Bouret, P., V. Rivoirard, and C. Tuleau-Malot. 2013. "Inference of Functional Connectivity in Neurosciences Via Hawkes Processes." In *2013 IEEE Global Conference on Signal and Information Processing* 317–320. Austin: IEEE.

Ridout, M. S., and M. Linkie. 2009. "Estimating Overlap of Daily Activity Patterns from Camera Trap Data." *Journal of Agricultural, Biological, and Environmental Statistics* 14: 322–337.

Rota, C. T., M. A. R. Ferreira, R. W. Kays, T. D. Forrester, E. L. Kalies, W. J. McShea, A. W. Parsons, and J. J. Millspaugh. 2016. "A Multispecies Occupancy Model for Two or More Interacting Species." *Methods in Ecology and Evolution* 7: 1164–73.

Say-Sallaz, E., S. Chamaillé-Jammes, H. Fritz, and M. Valeix. 2019. "Non-consumptive Effects of Predation in Large Terrestrial Mammals: Mapping our Knowledge and Revealing the Tip of the Iceberg." *Biological Conservation* 235: 36–52.

Schliep, E. M., A. E. Gelfand, J. S. Clark, and R. Kays. 2018. "Joint Temporal Point Pattern Models for Proximate Species Occurrence in a Fixed Area Using Camera Trap Data." *Journal of Agricultural, Biological and Environmental Statistics* 23: 334–357.

Schmitt, M. H., K. Stears, C. C. Wilmers, and A. M. Shrader. 2014. "Determining the Relative Importance of Dilution and Detection for Zebra Foraging in Mixed-Species Herds." *Animal Behaviour* 96: 151–58.

Searle, C. E., J. B. Smit, J. J. Cusack, P. Strampelli, A. Grau, L. Mkuburo, D. W. Macdonald, A. J. Loveridge, and A. J. Dickman. 2021. "Temporal Partitioning and Spatiotemporal Avoidance among Large Carnivores in a Human-Impacted African Landscape." *PLoS One* 16: e0256876.

Valiente-Banuet, A., M. A. Aizen, J. M. Alcántara, J. Arroyo, A. Cocucci, M. Galetti, M. B. García, et al. 2015. "Beyond Species Loss: The Extinction of Ecological Interactions in a Changing World." *Functional Ecology* 29: 299–307.

Zurell, D., U. Berger, J. S. Cabral, F. Jeltsch, C. N. Meynard, T. Münkemüller, N. Nehrbass, et al. 2010. "The Virtual Ecologist Approach: Simulating Data and Observers." *Oikos* 119: 622–635.

SUPPORTING INFORMATION

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